# APPENDIX B. CASE STUDY II: SUPPORTING MATERIALS

# **B.1. CASE STUDY II: MATRICES OF SCATTER PLOTS AND ABSOLUTE SPEARMAN CORRELATION COEFFICIENTS**



**Figure B-1.** Anions. Matrix of scatter plots and absolute Spearman correlation coefficients between specific conductivity ( $\mu$ S/cm), alkalinity (mg/L), sulfate (mg/L), chloride (mg/L), and ion ([HCO<sub>3</sub><sup>-</sup> + SO<sub>4</sub><sup>2-</sup>] mg/L) concentrations in Case Study II. All variables are logarithm transformed. The smooth lines are the locally weighted scatterplot smoothing (LOWESS) lines (span = 2/3).



**Figure B-2. Cations.** Matrix of scatter plots and absolute Spearman correlation coefficients between specific conductivity ( $\mu$ S/cm), ion ([HCO<sub>3</sub><sup>-</sup> + SO<sub>4</sub><sup>2-</sup>] mg/L), hardness (mg/L), Mg (mg/L), and Ca (mg/L), in Case Study II. All variables are logarithm transformed. The smooth lines are the locally weighted scatter plot smoothing (LOWESS) lines (span = 2/3).



**Figure B-3. Dissolved metals.** Matrix of scatter plots and absolute Spearman correlation coefficients among specific conductivity ( $\mu$ S/cm), ion ([HCO<sub>3</sub><sup>-</sup> + SO<sub>4</sub><sup>2-</sup>] mg/L), and dissolved metal concentrations (mg/L) in Case Study II. All variables are logarithm transformed. The smooth lines represent the locally weighted scatterplot smoothing (LOWESS) lines (span = 2/3).



**Figure B-4. Other water-quality parameters.** Matrix of scatter plots and absolute Spearman correlation coefficients between environmental variables in Case Study II. The smooth lines are locally weighted scatter plot smoothing (LOWESS) lines (span = 2/3). Specific conductivity is logarithm transformed specific conductivity ( $\mu$ S/cm); temp is water temperature (°C); Hab\_Sc is habitat score from Rapid Bioassessment (Habitat) Protocol (Barbour et al., 1999) score (possible range from 0 to 200); fecal is logarithm transformed fecal coliform bacteria count (per 100 mL water); embeddedness is a parameter score from the Rapid Bioassessment Protocol (possible range from 0 to 20); DO is dissolved oxygen (mg/L); TP is logarithm transformed total phosphorus (mg/L); NO23 is logarithm-transformed nitrite [NO<sub>2</sub><sup>-</sup>] plus nitrate [NO<sub>3</sub><sup>-</sup>] in mg/L.

#### **B.2. CASE STUDY II: ASSESSMENT OF POTENTIAL CONFOUNDERS**

Previous assessments of the factors potentially influencing the model of the causal relationship between ionic concentration and extirpation of benthic invertebrates (Suter and Cormier, 2013; U.S. EPA, 2011, Appendix B) indicated that the following factors did not substantially confound the causal relationship between specific conductivity (SC) and benthic macroinvertebrate assemblages: rapid bioassessment protocol (RBP) habitat scores (Barbour et al., 1999), sampling date, organic enrichment, nutrients, deposited sediments, high pH, selenium, heat (temperature), lack of headwaters, size of catchment area, settling ponds, dissolved oxygen (DO), and metals. However, low pH could possibly affect the model (Suter and Cormier, 2013; U.S. EPA, 2011, Appendix B) because its mode of action is associated with increased solubility of metals which are toxic (e.g., Wren and Stephenson, 1991; Ormerod et al., 1987). As a result, sampling sites with acidic waters (pH <6) were excluded from the analysis in order to minimize any effects, but no other modification of the data set was required to address confounding.

New analyses described below are consistent with the analyses reported by the (U.S. EPA, 2011).

#### **B.2.1.** Multivariate Analysis

Potential confounding of the model for Case Study II was reassessed for habitat (total RBP score), embeddedness (RBP subscore), temperature, and organic enrichment (fecal coliform) using a two-step process.

Habitat quality and fecal coliform together had little effect on the slope in multiple regression analyses with the dependent variable of occurrence of the genera with the 36 lowest extirpation concentration (XC<sub>95</sub>) values (see Table B-1). However, to ensure that they were not influential, their combined effect on the hazardous concentration (HC<sub>05</sub>) was determined (see Section B.2.2). The most influential parameter other than SC was temperature (slope = -0.202, Spearman standard error [SE] 0.21). Because there is a relationship with the life history of salt-intolerant taxa and because there is a nonlinear relationship between temperature and sampling date (see Figure B-5), further analyses were performed to evaluate temperature/sampling date related to the HC<sub>05</sub> (see Section B.2.3).

**Table B-1. An output table for two generalized linear models.** The first is the simple model predicting the number of mayfly genera from specific conductivity. The second is a multivariate model with the additional covariates rapid bioassessment protocol (RBP) score, temperature, and fecal coliform count. These variables were chosen based on previous analyses as likely confounders that could co-occur and have combined effects. Fecal coliform count and specific conductivity were first log10 transformed to normalize the data, then all four variables were centered and scaled (subtracting the means and then dividing the centered values by their standard deviation) so that all four variables are at the same scales. The response variable is assumed to follow a Poisson distribution which appropriate for counts of occurrences.

Parameter	Estimate	Standard error
Univariate model		
Intercept	0.848	0.017
Specific conductivity slope	-0.852	0.018
Multivariate model		
Intercept	0.842	0.017
Specific conductivity slope	-0.703	0.019
RBP slope	0.037	0.013
Temperature slope	-0.202	0.013
Fecal coliform slope	-0.077	0.013



Figure B-5. Scatter plot showing inter-relatedness of stream temperature and sampling date. The fitted line is a locally weighted scatterplot smoothing spline (LOWESS, quadratic polynomial, span = 0.75).

### **B.2.2.** Influence of Poor Habitat and Organic Enrichment on the Hazardous Concentration (HC<sub>05</sub>)

To assure that the genus extirpation concentration distribution (XCD) model was detecting effects from SC and not a response to poor habitat, the HC<sub>05</sub> was recalculated using the example criterion-data set in which samples were removed with an RBP score <130 total, the HC<sub>05</sub> was 337  $\mu$ S/cm. The threshold of RBP <130 was selected as an upper bound on acceptable habitat by Gerritsen et al. (2010) that also provided an adequate and to maximize sample size (relevant *n* = 581). This threshold of RBP <130 represents, on average, habitat that is not pristine, but which is adequate for maintenance of biological assemblages. Removal of poor habitat and high fecal coliform samples from the data set had almost no effect on the XCD model

or HC<sub>05</sub> (see Figure B-6). With this constrained data set (RBP score >130) the HC<sub>05</sub> was 337  $\mu$ S/cm (95% confidence interval [CI] 265–360  $\mu$ S/cm). The confidence interval overlaps with the HC<sub>05</sub> for the example criterion continuous concentration (CCC; 338  $\mu$ S/cm; 95% CI 272–365  $\mu$ S/cm). Therefore, no correction was made for habitat quality or organic enrichment.





#### **B.2.3.** Potential Influence of Temperature on the Hazardous Concentration (HC<sub>05</sub>)

To assure that the genus XCD model was detecting effects from SC and not a response to warmer temperatures, the example criterion data set was constrained to samples with pH <6 and fecal coliform <400 colonies/100 mL and either a temperature  $\geq 17^{\circ}$ C or <17°C. The threshold of 17°C was selected based on reported upper temperature tolerance values for aquatic insects (Nebeker and Lemke, 1968; Vieira et al., 2006) and provided adequate sample sizes. If low temperature is a confounder, the XCD <17°C is expected to move to the right because lower temperatures are less stressful and organisms may be able to tolerate higher SC levels. If high temperature is a confounder, conditions are more stressful and the XCD  $\geq 17^{\circ}$ C is expected move to the left; that is, lower XC<sub>95</sub> values and a lower HC<sub>05</sub>.

Removal of cooler samples from the data set decreased the HC<sub>05</sub> (see Figure B-7). With the data set constrained to temperatures  $\geq 17^{\circ}$ C, the HC<sub>05</sub> was 315 µS/cm (95% CI = 256–365, 116 genera, relevant n = 1,416 samples). This is consistent with confounding by higher temperatures. Removal of warmer samples from the data set increased the HC<sub>05</sub> (see Figure B-7). With the data set constrained to temperatures <17°C, the HC<sub>05</sub> was 425 µS/cm (95% CI = 292–455, 95 genera, relevant n = 658 samples). This is also consistent with the direction that is expected to occur if temperature is a confounder. Hence, the results are logically inconsistent with temperature acting as a cause of extirpation along with SC. The confidence interval overlaps with the unconstrained example HC<sub>05</sub> (338 µS/cm; 95% CI 272–365 µS/cm, 139 genera). Furthermore, the reduced sample sizes reduced the overall number of genera in the model by 17–32%. Also, the XCD overlaps the full data set through most of the lower range of the models. In such cases, correction for confounding may increase error in the estimated HC<sub>05</sub>, therefore, no correction was made for temperature.



Figure B-7. Genus extirpation concentration (XC<sub>95</sub>) distributions for example criterion data set and temperature constrained data sets. Samples with pH  $\leq 6$  and (SO<sub>4</sub> + HCO<sub>3</sub>)/Cl  $\leq 1$  were removed from all data sets. The example criterion (unconstrained, 0–32°C) data set has XC<sub>95</sub> values for 139 genera (open black diamonds). The  $\leq 17^{\circ}$ C constrained data set has 95 genera (closed green diamonds [N = 658]). The  $\geq 17^{\circ}$ C constrained data set has 116 genera (open red triangles [n = 1,416]). For comparison the 5<sup>th</sup> centile is 338 µS/cm in the unconstrained data set, 425 µS/cm in the  $\leq 17^{\circ}$ C constrained data, and 315 µS/cm in the  $\geq 17^{\circ}$ C constrained data.

#### B.2.4. Potential Influence of Sampling Date on the Hazardous Concentration (HC<sub>05</sub>)

To assess effects of date of sampling on the XCD model, three lines of evidence were analyzed to address potential confounding by lack of seasonal observation of apparently salt-intolerant genera. First, the  $HC_{05}$  using the spring (March to June) only data set was compared with the full-year data set. (Seasons are defined by the phenology of the aquatic insects and the changes in SC, not the conventional intervals.) The confidence bounds of the spring  $HC_{05}$  overlap with the confidence bounds of the full-year data set (see Figure B-8). The summer (July to October) only XCD lacks taxa known to be intolerant to ionic stress which can be seen by the overall shift of the XCD to the right. The shift in the upper portion of the XCD in spring is mostly due to the narrower SC sample range during the spring compared to the all year data set.

Next, a scatter plot and regression model was developed for the relationship between measurements of SC at the time of the biological sample and annual mean SC (see Figure B-9). The annual geometric mean SC values were calculated from at least six water samples collected before biological samples were taken. At least one spring and one summer sample were required in order to be included in the data set. There were 325 sites with paired SC and biological data (see Figure B-9) meeting these additional data requirements. On the *x*-axis is the SC value when biological samples were collected and on the *y*-axis is the annual geometric mean value during that rotating year for a site. A Model II Regression was fitted for this data set which takes into account for error variance in both variables. The mean relationship between measurements of SC at the time of the biological sample and annual mean SC is nearly 1:1. For example, when SC is 340  $\mu$ S/cm on the biological sampling date, the regression prediction for an annual mean SC for the same site is 304  $\mu$ S/cm.



Figure B-8. Comparison of genus extirpation concentration distributions (XCD) full data set and subsets in different months. Example criterion data set (black circles) and subsets of March to June (spring, inverted green triangles) and July to October (summer, filled red triangles) collected samples from the Case Study II Criterion-data set. The all year XCD has 139 genera, the spring XCD has 110 genera (N = 1,044), summer XCD has 92 genera (N = 989). The horizontal dotted line is the 5<sup>th</sup> centile. The spring and summer hazardous concentrations' (HC<sub>05</sub>) 95% confidence bounds overlap with the all-year data set. The summer XCD model lacks salt-intolerant genera in the data set.



Model II Regression and 95% confidence Intervals

Simultaneous biosample specific conductivity



Lastly, to account for the seasonal variability, SC values collected at the time of biological sampling were adjusted to estimate annual mean SC values as described in Section 3.1.4. The weighting factors vary slightly for different months 0.94 to 1.05 (see Table B-2). June through November SC values are slightly higher than the annual average, so the weighting factors are generally lower, while the earlier spring weighting factors are generally higher. The HC<sub>05</sub> calculated with weighted SC measurements is 385  $\mu$ S/cm (CI 327–468  $\mu$ S/cm;

see Figure B-10). These three analyses suggest that sampling date is at most a minor confounder. Correction for confounding may increase error in the estimated  $HC_{05}$ , therefore, no correction was made for sampling date.

 Table B-2. Weighting factors used to normalize specific conductivity on date

 of biological sample to annual average

Month	1	2	3	4	5	6	7	8	9	10	11	12
Weighting factor	1.03	1.05	1.04	1.05	1.05	0.99	0.96	0.95	0.94	0.94	0.97	1.00



Figure B-10. Case Study II comparison of annual weighted and original extirpation concentrations (XC95). Genus extirpation concentration distribution (XCD) of unweighted XC95 values (gray) and XCD of XC95 derived from specific conductivity normalized to an annual geometric mean (blue). Hazardous concentration (HC05) values are 338, and 385  $\mu$ S/cm, respectively.

#### **B.2.5.** Conclusion

Previous assessments of the factors potentially influencing the model of the causal relationship between ionic concentration and extirpation of benthic invertebrates indicated that 13 factors had little or no effect on the causal relationship between SC and benthic macroinvertebrate assemblages (Suter and Cormier, 2013; U.S. EPA, 2011, Appendix B). The 13 factors that were considered included RBP habitat scores, sampling date, organic enrichment, nutrients, deposited sediments, high pH, selenium, heat (temperature), lack of headwaters, size of catchment area, settling ponds, dissolved oxygen, and metals.

The additional analyses described in this Appendix Section B.2 using data from Ecoregion 70 indicate that SC remains the strongest influence in the multivariate model of genera with low XC<sub>95</sub> values (see Table B-1). Organic enrichment (estimated based on fecal coliform counts) did not significantly contribute to the multivariate model and no further analyses were warranted. Habitat score showed a minor effect in the multivariate model, but recalculating the HC<sub>05</sub> in a data set with sites with RBP score >130 resulted in an HC<sub>05</sub> with confidence intervals that overlapped with the HC<sub>05</sub> from the example criterion data set. Temperature and sample date are nonlinearly associated; therefore, three different analyses were performed to assess potential effects on the XCD model. They indicated that neither temperature nor sample date confounds the HC<sub>05</sub> of the XCD model for the example criteria. Therefore, the example criterion data set was not altered and no corrections were made for habitat, temperature, or sample date.

# **B.3. CASE STUDY II: COMPARISON OF WATER CHEMISTRY BASED CRITERION MAXIMUM EXPOSURE CONCENTRATION (CMEC) AND BIOLOGICAL SURVIVAL**

The criterion maximum exposure concentration (CMEC) is the maximum SC level that may occur for a short duration and be protective of 95% of macroinvertebrate genera. The CMEC for Ecoregion 70 was calculated using the water chemistry approach in Section 3.2. In this method, the CMEC is estimated at the 90<sup>th</sup> centile of observations at sites with water chemistry regimes meeting the CCC. It estimates the protective maximum using only water chemistry data without biological data.

Owing to the moderate number of biological samples with multiple seasonal sampling of water chemistry available for Ecoregion 70, it was possible to estimate a maximum SC that could

occur and salt-intolerant genera had survived until the following year shortly before emergence as winged adults. Salt-intolerant genera are more commonly observed when they are larger and nearing emergence usually in April through June. The maximum SC of streams in Ecoregion 70 usually occurs between August and September.

A data set was constructed from the Ecoregion 70 criterion-data set. For a site to be included, it required a minimum of six water chemistry samples taken samples of water chemistry data taken over the course of the year prior to biological sampling. A minimum of six samples were required for inclusion in the data set which was defined as a rotating year. Of the 819 sites sampled in the data set, 317 met the stringent criteria for inclusion in the data set for this analysis. The data set tended to represent long term reference sites and sites monitored for remediation. Therefore, the data set is not optimal for this analysis. However, it is useful for illustrating the analysis and for evaluating the degree the protectiveness of the CMEC estimated from SC measurements alone.

The relationship between SC and the presence of salt-intolerant taxa were inspected for each of the 317 sites that met the inclusion criteria in the data set. The most salt-intolerant taxa are those taxa with and  $XC_{95} <340 \ \mu$ S/cm, the CCC for Ecoregion 70. Figure B-11 depicts two plots where the annual average SC is well below 340  $\mu$ S/cm. Fivemile Creek is an exceptional site with three of the seven salt-intolerant taxa, moderate temperature with some higher levels in summer months, and low SC year-round with an annual average of 240  $\mu$ S/cm and a maximum of 487  $\mu$ S/cm. Fivemile Creek has four salt-intolerant taxa with an annual average of 158.5  $\mu$ S/cm and a maximum grab sample of 460  $\mu$ S/cm. Buffalo Creek would meet the recommended SC CCC and for Ecoregion 70.



Figure B-11. Specific conductivity and temperature variations in stations with multiple observations. Julian day, 0 = January, is on the *x*-axis. Specific conductivity is on the left *y*-axis with water chemistry samples as open circles; a filled circle is date of biological sampling. Dashed line is at 340 µS/cm for orientation. One rotating year is defined as the year prior to biological sample, a minimum of six samples were required for inclusion in the data set. Specific conductivity minimum (min), maximum (max), and date of biological sampling (bio) are shown in the lower left corner. The count of the seven most salt-intolerant genera (extirpation concentration [XC<sub>95</sub>] <340 µS/cm) are 4 for Little Buffalo and 3 Fivemile Creeks.

As an evaluation of the CMEC, an analysis was performed to compare the calculated CMEC with an estimate of a tolerated maximum SC using biological survival as the assessment endpoint. A scatter plot was constructed of the count of the seven most salt-intolerant taxa and maximum SC that occurred in the year prior to biological sampling (see Figure B-12). The analysis showed that there is a negative relationship between maximum SC and salt-intolerant genera. There are few observations of salt-intolerant genera were observed at sites with SC >680  $\mu$ S/cm in this data set, the CMEC calculated from chemistry only data. The chemistry only analysis used a much more representative sample of sites comprised of 819 rotation years from 805 unique stations, with at least one sample from July to October (J–O) and one from March to June (M–J), and at least six samples within a rotation year (see Table 5-3). Because the CMEC

analysis is based on a much larger and more representative sample, the CMEC of 680  $\mu$ S/cm was retained. However, as data becomes more available, the method using biological samples may become preferable.



Figure B-12. Scatter plot of count of salt-intolerant genera (extirpation concentration [XC95] <340  $\mu$ S/cm) and maximum conductivity in preceding year. Few salt-intolerant genera are observed at sites (N = 317) with a specific conductivity greater than the criterion maximum exposure concentration (CMEC) of 680  $\mu$ S/cm (vertical dashed line). Specific conductivity expressed as  $\mu$ S/cm.

Order	Family	Genus	Symbol	XC95	95% CI	N
Ephemeroptera	Ephemerellidae	Drunella		136	127-169	78
Plecoptera	Chloroperlidae	Utaperla		248	193-323	31
Ephemeroptera	Heptageniidae	Cinygmula		258	207-329	120
Plecoptera	Chloroperlidae	Alloperla		275	229-307	71
Ephemeroptera	Ephemerellidae	Ephemerella		283	237-364	316
Ephemeroptera	Heptageniidae	Heptagenia		294	229-563	41
Plecoptera	Perlodidae	Diploperla		338	257-404	152
Ephemeroptera	Ephemerellidae	Eurylophella		346	270-450	146
Ephemeroptera	Heptageniidae	Nixe		359	314-560	200
Diptera	Dixidae	Dixa		398	325-1,247	25
Plecoptera	Chloroperlidae	Haploperla		403	341-566	212
Plecoptera	Perlodidae	Isoperla		460	377-567	541
Trichoptera	Glossosomatidae	Agapetus		466	287-502	25
Ephemeroptera	Heptageniidae	Epeorus		481	359-2,020	303
Trichoptera	Uenoidae	Neophylax		499	317-577	144
Diptera	Ceratopogonidae	Bezzia		514	281-563	37
Ephemeroptera	Ameletidae	Ameletus		567	272-4,884	244
Diptera	Chironomidae	Demicryptochironomus		618	297-857	66
Diptera	Chironomidae	Zavrelia		627	275-1,383	60
Diptera	Chironomidae	Conchapelopia		640	393-1,175	121
Plecoptera	Perlidae	Eccoptura		648	440-1,028	51
Ephemeroptera	Baetidae	Plauditus		688	567-756	365
Ephemeroptera	Baetidae	Diphetor		701	565-951	133
Ephemeroptera	Leptophlebiidae	Paraleptophlebia		706	508-812	400
Ephemeroptera	Baetidae	Procloeon		708	646-1,252	87
Ephemeroptera	Heptageniidae	Leucrocuta		727	358-1,082	217
Ephemeroptera	Baetiscidae	Baetisca		757	264-762	34
Ephemeroptera	Heptageniidae	Maccaffertium		783	672-1,017	440
Trichoptera	Limnephilidae	Pycnopsyche	~	784	456-1,228	26
Ephemeroptera	Leptophlebiidae	Leptophlebia		805	277-912	59
Coleoptera	Dytiscidae	Hydroporus		822	347-822	37
Plecoptera	Peltoperlidae	Peltoperla	~	824	379-1,175	32
Plecoptera	Nemouridae	Amphinemura		911	531-3,725	618
Diptera	Chironomidae	Potthastia		944	480-1,059	32
Ephemeroptera	Heptageniidae	Stenonema		945	653-1,075	614

# B.4. CASE STUDY II: EXTIRPATION CONCENTRATION (XC95) VALUES

Order	Family	Genus	Symbol	XC95	95% CI	N
Trichoptera	Philopotamidae	Wormaldia	~	947	459-1,261	35
Ephemeroptera	Baetidae	Acentrella		986	505-3,162	567
Isopoda	Asellidae	Asellus		1,014	365-1,014	26
Ephemeroptera	Isonychiidae	Isonychia		1,017	805-1,129	654
Plecoptera	Perlodidae	Cultus	~	1,073	307-1,398	27
Diptera	Chironomidae	Stempellinella		1,077	951-1,338	262
Odonata	Gomphidae	Lanthus	~	1,091	566-1,175	29
Ephemeroptera	Heptageniidae	Stenacron	~	1,100	973-1,195	245
Ephemeroptera	Baetidae	Centroptilum	~	1,137	508-1,195	72
Isopoda	Asellidae	Lirceus		1,247	566-1,534	131
Decapoda	Cambaridae	Cambarus	>	1,278	1,046-1,540	307
Diptera	Chironomidae	Parachaetocladius	>	1,285	1,166-1,665	40
Diptera	Chironomidae	Brillia	>	1,301	582-1,526	45
Coleoptera	Psephenidae	Ectopria	>	1,346	978-2,148	214
Veneroida	Pisidiidae	Pisidium	>	1,402	1,287-1,470	49
Diptera	Chironomidae	Parakiefferiella	>	1,569	1,419–1,700	52
Coleoptera	Elmidae	Macronychus	>	1,605	1,195–1,678	63
Ephemeroptera	Baetidae	Baetis		1,620	1,197–2,580	1,222
Diptera	Ceratopogonidae	Dasyhelea	>	1,696	1,136-1,864	48
Diptera	Chironomidae	Natarsia	>	1,786	1,613-1,842	48
Diptera	Empididae	Chelifera	>	1,845	1,589-1,870	39
Diptera	Chironomidae	Cardiocladius	>	1,951	1,270-1,951	120
Diptera	Chironomidae	Pagastia	>	1,970	1,480-1,970	38
Diptera	Chironomidae	Eukiefferiella	>	1,977	1,598–2,523	305
Trichoptera	Rhyacophilidae	Rhyacophila	~	1,977	631-5,057	191
Amphipoda	Crangonyctidae	Crangonyx	>	1,978	734-1,978	111
Ephemeroptera	Ephemeridae	Ephemera		1,978	475-1,978	90
Hemiptera	Veliidae	Rhagovelia	>	2,030	1,171-2,030	27
Diptera	Simuliidae	Prosimulium	>	2,148	550-2,439	141
Plecoptera	Leuctridae	Leuctra	>	2,257	1,523-2,791	1,010
Diptera	Chironomidae	Sublettea	>	2,294	1,367–2,294	124
Diptera	Chironomidae	Chaetocladius	>	2,320	1,700-5,057	170
Diptera	Chironomidae	Krenopelopia	>	2,320	1,700-2,320	44
Diptera	Chironomidae	Phaenopsectra	>	2,332	1,348-2,332	61
Diptera	Tipulidae	Tipula	>	2,420	1,902-6,492	532
Hemiptera	Veliidae	Microvelia	>	2,523	978-2,523	31

Order	Family	Genus	Symbol	XC95	95% CI	N
Diptera	Chironomidae	Orthocladius		2,523	500-2,523	167
Diptera	Chironomidae	Rheopelopia		2,523	410-2,523	72
Coleoptera	Elmidae	Microcylloepus	>	2,558	1,397–2,558	76
Diptera	Chironomidae	Tvetenia	>	2,573	1,729–2,791	516
Diptera	Chironomidae	Nilotanypus	>	2,630	1,422-2,630	119
Trichoptera	Polycentropodidae	Polycentropus	>	2,641	1,410-4,713	192
Trichoptera	Hydroptilidae	Ochrotrichia	>	2,791	1,143-2,791	48
Decapoda	Cambaridae	Orconectes	>	3,162	1,520-3,162	211
Diptera	Chironomidae	Paratanytarsus	>	3,489	3,162-5,258	102
Ephemeroptera	Caenidae	Caenis	>	3,884	2,641-4,052	693
Diptera	Chironomidae	Microtendipes	>	3,972	2,437-7,053	462
Diptera	Chironomidae	Rheotanytarsus	>	4,400	2,605-5,468	875
Diptera	Chironomidae	Corynoneura	>	4,636	980-4,636	125
Diptera	Chironomidae	Diamesa	>	4,636	1,924–4,713	463
Diptera	Chironomidae	Polypedilum	>	4,636	3,314-7,093	1,598
Diptera	Chironomidae	Rheocricotopus	>	4,636	1,902-4,884	533
Megaloptera	Sialidae	Sialis	>	4,636	3,714-11,227	241
Isopoda	Asellidae	Caecidotea	>	4,713	1,977-4,713	178
Diptera	Empididae	Clinocera	>	4,713	577-4,713	36
Amphipoda	Gammaridae	Gammarus	>	4,713	2,320-10,350	232
Diptera	Chironomidae	Parametriocnemus	>	4,713	2,580-4,884	1,266
Diptera	Chironomidae	Zavrelimyia	>	4,884	1,589–4,884	212
Coleoptera	Elmidae	Oulimnius	>	5,000	824-5,000	30
Diptera	Tipulidae	Limonia	>	5,057	1,687-5,057	35
Diptera	Chironomidae	Limnophyes	>	5,120	1,445-5,120	48
Diptera	Chironomidae	Cryptochironomus	>	5,258	2,580-7,093	356
Diptera	Chironomidae	Larsia	>	5,258	875-5,258	98
Diptera	Chironomidae	Tribelos	>	5,258	1,081-5,258	55
Diptera	Tipulidae	Antocha	>	6,468	3,972-7,093	433
Diptera	Chironomidae	Micropsectra	>	6,468	2,471-6,468	173
Diptera	Chironomidae	Paraphaenocladius	>	6,468	863-6,468	43
Diptera	Simuliidae	Simulium	>	6,468	2,874-7,053	1,001
Odonata	Gomphidae	Stylogomphus	>	6,468	2,320-6,468	70
Trichoptera	Hydropsychidae	Ceratopsyche	>	6,492	4,713-7,010	745
Trichoptera	Philopotamidae	Chimarra	>	6,492	3,489-7,053	587
Trichoptera	Hydropsychidae	Diplectrona	>	6,492	1,870-6,492	277

Order	Family	Genus	Symbol	XC95	95% CI	N
Plecoptera	Perlidae	Perlesta		6,492	1,139–6,492	453
Diptera	Tipulidae	Pseudolimnophila	>	6,492	1,589–6,492	142
Plecoptera	Chloroperlidae	Sweltsa	~	6,492	916-6,492	256
Diptera	Chironomidae	Thienemannimyia	>	6,492	4,400-7,093	1,361
Coleoptera	Elmidae	Stenelmis	>	6,620	3,972-7,370	1,705
Trichoptera	Philopotamidae	Dolophilodes	>	7,053	588-7,053	85
Coleoptera	Psephenidae	Psephenus	>	7,105	4,884-7,370	1,000
Odonata	Aeshnidae	Boyeria	>	7,340	2,407-7,340	133
Megaloptera	Corydalidae	Nigronia	>	7,340	3,162-9,790	503
Basommatophora	Physidae	Physella	>	7,340	6,468-9,790	183
Diptera	Tipulidae	Dicranota	>	7,370	2,145-7,370	233
Coleoptera	Elmidae	Dubiraphia	>	7,370	3,162-7,370	200
Diptera	Tipulidae	Hexatoma	>	7,370	6,468-9,790	811
Coleoptera	Elmidae	Optioservus	>	7,370	4,713-9,790	1,231
Trichoptera	Hydropsychidae	Cheumatopsyche	>	9,180	5,266-9,180	1,569
Diptera	Chironomidae	Tanytarsus	>	9,180	4,636-9,790	1,183
Diptera	Tabanidae	Tabanus	>	9,790	2,291-9,790	72
Diptera	Chironomidae	Thienemanniella	>	9,790	6,573-11,227	364
Trichoptera	Hydropsychidae	Hydropsyche	>	10,140	4,884-10,140	769
Diptera	Empididae	Hemerodromia	>	10,350	7,010-11,646	427
Megaloptera	Corydalidae	Corydalus	>	11,227	7,340-11,646	257
Diptera	Chironomidae	Cricotopus	>	11,227	6,468-11,646	504
Diptera	Chironomidae	Dicrotendipes	>	11,227	9,790-11,646	313
Trichoptera	Hydroptilidae	Hydroptila	>	11,227	4,884-11,646	386
Diptera	Chironomidae	Procladius	>	11,227	2,630-11,227	35
Diptera	Chironomidae	Ablabesmyia	>	11,582	7,370-11,646	184
Plecoptera	Perlidae	Acroneuria	>	11,646	1,066-11,646	287
Diptera	Athericidae	Atherix	>	11,646	7,340-11,646	80
Diptera	Tabanidae	Chrysops	>	11,646	7,053-11,646	70
Diptera	Chironomidae	Cladotanytarsus	>	11,646	5,253-11,646	121
Coleoptera	Dryopidae	Helichus	>	11,646	1,270-11,646	325
Diptera	Chironomidae	Pseudochironomus	>	11,646	4,400-11,646	50

# B.5. CASE STUDY II: GENERALIZED ADDITIVE MODEL (GAM) PLOTS

The generalized additive model (GAM) plots in this Appendix Section B.5 were used to designate ~ and > values for those  $XC_{95}$  values listed in Appendix Section B.4. In this example, the probability of observing a genus is the proportion of sampled stations in a conductivity bin with the genus present based on taxonomic identification of 200 individuals per sample. Conductivity is reported as specific conductivity. The red, dashed vertical line is the  $XC_{95}$  value for the genus (see Section B.4) obtained from the plots of the cumulative distribution function's (CDFs) in Appendix Section B.6. Plots are arranged from the lowest to the highest  $XC_{95}$  value.





























B-34



# B.6. CASE STUDY II: CUMULATIVE DISTRIBUTION FUNCTION (CDF) PLOTS

The CDFs used to derive the  $XC_{95}$  values are shown in this Appendix Section B.6. Conductivity is reported as specific conductivity. The red, dashed vertical line is the  $XC_{95}$  value for the genus (see Section B.4) obtained from each plotted CDF in Appendix Section B.6. Plots are arranged from the lowest to the highest  $XC_{95}$  value.



**B-37** 



**B-38** 















B-43



**B-44** 











**B-47** 





.....

m

0.2

8

тт

## **B.7. REFERENCES**

Barbour, MT; Gerritsen, J; Snyder, BD; Stribling, JB. (1999) Rapid bioassessment protocols for use in streams and wadeable rivers: periphyton, benthic macroinvertebrates, and fish. 2<sup>nd</sup> ed. EPA/841/B 99/002. Washington, DC: U.S. Environmental Protection Agency, Office of Water. Available online at <a href="http://www.epa.gov/owow/monitoring/rbp/wp61pdf/rbp.pdf">http://www.epa.gov/owow/monitoring/rbp/wp61pdf/rbp.pdf</a>.

Gerritsen, J; Zheng, L; Burton, J; Boschen, C; Wilkes, S; Ludwig, J; Cormier, S. (2010) Inferring causes of biological impairment in the Clear Fork Watershed, West Virginia. EPA/600/R-08/146. Cincinnati, OH: U.S. Environmental Protection Agency, Office of Research and Development, National Center for Environmental Assessment. <u>http://oaspub.epa.gov/eims/eimscomm.getfile?p\_download\_id=496962</u>.

Nebeker AV; Lemke, AE. (1968) Preliminary studies on the tolerance of aquatic insects to heated waters. J Kansas Entomol Soc. 41(3):413–418.

Ormerod, SJ; Boole, P; McCahon, CP; Weatherley, NS; Pascoe, D; Edwards, RW. (1987) Short-term experimental acidification of a Welsh stream: comparing the biological effects of hydrogen ions and aluminum. Freshw Biol 17: 341–356.

Suter, GW; Cormier, SM. (2013) A method for assessing the potential for confounding applied to ionic concentration in central Appalachian streams. Environ Toxicol Chem 32: 288–295. http://dx.doi.org/10.1002/etc.2054.

U.S. EPA (U.S. Environmental Protection Agency). (2011) A field based aquatic life benchmark for conductivity in Central Appalachian streams. EPA/600/R-10/023F. Washington, DC: Office of Research and Development, National Center for Environmental Assessment. <u>http://cfpub.epa.gov/ncea/cfm/recordisplay.cfm?deid=233809</u>.

Vieira, NKM; Poff, NL; Carlisle, DM; Moulton, SR, II; Koski, ML; Kondratieff, BC. (2006) A database of lotic invertebrate traits for North America: U.S. Geological Survey Data Series 187. U.S. Department of the Interior, U.S. Geological Survey. Available online at http://pubs.water.usgs.gov/ds187.

Wren, CD; Stephenson, GL. (1991) The effect of acidification on the accumulation and toxicity of metals to freshwater invertebrates. Environ Pollut 71: 205–241.